INFLUENCE OF ACETATE AND FORMATE BASED DEICERS ON ASR IN AIRFIELD CONCRETE PAVEMENTS

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Abstract

In the past few years, the issue of external alkalis and their influence on ASR in concrete has become more important since several airfield concrete pavements showed ASR-distress related to the use of alkali-containing runway deicers based on acetates and formates. To assess the influence of external alkalis on ASR in pavement concretes, the cyclic climate storage has revealed as suitable performance-test. Especially deicers based on acetates and formates turned out to be extremely deleterious in concrete with reactive aggregates. Experiments with model pore solutions and cement pastes, as well as speciation calculations were performed to investigate possible mechanisms. The obtained results indicate that the solubility of portlandite in the cement paste is increased in presences of acetate and formate based deicers. Due to the additional release of OH⁻ from portlandite and the supply of alkalis, ASR can be initiated and accelerated highly in concretes with reactive aggregates. There are also indications for an interaction of other cement paste constituents with acetate and formate based deicers.

Keywords: airfield pavement concrete, deicer, acetate, formate, external alkali mechanism

1 INTRODUCTION

It has been suspected for some time now, that the action of alkali-containing deicers is capable of favoring an alkali-silica reaction (ASR) in concrete with reactive aggregates. The research in the past decades was mainly focused on NaCl [1, 2, 3]. For safety reasons, chloride-based deicers are not used on airfields in order to avoid corrosion damage to aircraft and airport equipment. Therefore, synthetic urea (CH₄N₂O) was used as runway deicer for a long time. Deicers based on sodium and potassium acetate (CH₃COOK) and later on also based on sodium and potassium formate (HCOOK), were introduced in the USA as well as in Germany in the early 1990s. These deicers exhibit a large number of advantages over the synthetic urea. They are more environmentally friendly, since they are readily biodegradable, leaving behind only harmless, non-toxic residues and contain no nitrogen as well. They are also more effective at lower temperatures, work faster and longer and require fewer applications. In a while, acetate based deicers are also used for some highways and bridge decks, to benefit from the environmental advantages and prevent reinforced concrete structures from chloride-induced corrosion damage.

However, in the past few years severe ASR-distress occurred on airfield concrete pavements with increasing frequency. It is reported that about thirty U.S. military airfields around the world and at least eight commercial U.S. airports are affected [4, 5]. At the FIB (F.A. Finger-Institut für Baustoffkunde), a German runway was investigated that showed ASR-distress 5-7 years after the deicer was changed from urea to potassium acetate. Commercial deicer solutions contain about 50-55 % active components and minor amounts of corrosion inhibitors. To improve the effectiveness, solid deicers (e.g. sodium formate) and liquid deicers (e.g. potassium formate) are often used together. The application rates depend strongly on the weather situation and are usually about 10-60 g/m². In Germany, there are approximately 40-50 days of deicing within every winter period and, if necessary, up to eight applications are done per day. So far, only a few studies became available that investigated the influence of these deicers on ASR in concrete.

Since 2001, an alternating climate test method (cyclic climate storage) is used at the FIB for accelerated simulation of Central European climatic conditions, in order to assess the durability of specific concretes for outdoor structures. Significant factors from environmental effects (drying, moistening, freezing and thawing, exposure to deicers) are simulated for this purpose by alternating temperature and moisture conditions in a special developed climate simulation chamber. A climate test program was especially developed for investigations into ASR for pavement concretes [6, 7]. After

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testing more than a hundred concretes, mostly job mixtures, with different cements and mostly with slow/late reacting aggregates, the results show clearly that a deleterious ASR can be initiated and highly accelerated in concretes with reactive aggregates exposed to alkali-containing deicers (Figure 1). It was found that externally supplied alkalis exceed the influence of the cement alkalis sooner or later. Low-alkali cements may delay the ASR, but are not capable of permanently preventing a deleterious ASR, if external alkalis are supplied and reactive aggregates are present [8]. To provide an explanation for these observations, investigations into the mechanisms were performed and are introduced below.

2 MATERIALS AND METHODS

2.1 Materials and experimental settings

Solubility experiments were performed using pure powders of $Ca(OH)_2$ (portlandite), CH_3COOK (K-acetate), HCOOK (K-formate), $Ca(CH_3COO)_2$ (Ca-diacetate) and $CaSO_4 \cdot 2H_2O$ (gypsum) for comparison purposes. The used water was double distilled and freshly boiled to remove dissolved carbon dioxide. The deicing chemicals were mixed in different concentrations (0.05, 0.3, 1.5, 2.5 mol/l) into saturated solutions (200 ml) of portlandite with solid portlandite in excess. The experiments were done in a glove-box, using nitrogen as inert gas to avoid any possible degradation of the acetate and formate ions in contact with oxygen and to avoid precipitation of calcium carbonate as well. Furthermore, in a saturated solution (200 ml) of gypsum with solid gypsum in excess, K-acetate (0.05, 0.3, 1.5, 2.5 mol/l) was mixed. All solutions were sealed airtight in plastic vials and stored overnight at 20 °C. The solutions were analyzed by ICP-OES and the pH was measured by using an alkali-resistant glass electrode (Orion ROSS, 81-08). Selected solid residues were filtered, dried at 40 °C and analyzed by XRD, FT-IR and ESEM/EDS.

Cement pastes were made from two portland cements with different Na₂O_{eq} (0.44 and 0.81 wt%) and w/c of 0.42. After 7 days curing in airtight sealed vials at 20 °C, the hardened samples (\emptyset 30 mm, l=30 mm) were cyclical, every 14 days, vacuum-immersed in two commercial runway deicer solutions (K-acetate, K-formate) as well as in water for control. The vacuum immersion was done 5-times in series of decompression (15 minutes) and compression. After the vacuum immersion, the samples were sealed airtight in a plastic bag and stored at 40 °C for 14 days. Afterwards, the pore solution was squeezed out and analyzed by ICP-OES and cement paste samples were also examined by XRD and ESEM/EDS.

2.2 Methods for assessment and analysis

Inductively coupled plasma-optical emission spectrometry (ICP-OES)

About 20 ml of the solutions were filtered (0.45 μ m) and sealed airtight in plastic vials. The solutions then were diluted (1:100) with nitric acid (1%) and about 100 μ l were injected in the argonplasma of the spectrometer (Perkin-Elmer, Optima 3000). The element concentrations of K, Na, Ca and S were determined. The detection limits ranging from 0.02 μ g/l (Ca) to 30 μ g/l (S).

X-ray powder diffraction (XRD)

Solid residues were prepared in the glove-box by separation from the solution with a frit (40 μ m) and rinse them with isopropanol and acetone. The residue was applied on a polyethylene sample holder, covered with a sheet of polypropylene and then discharged from the glove-box.

The cement paste samples were crushed dryly by hand with hammer and chisel and further on gently with a corundum mortar and pestle to obtain a coarse powder. The powder was rinsed with isopropanol and acetone in order to stop the hydration. Afterwards, the powder was dried for 30 minutes at 40 °C prior to gently grinding by hand with a mortar and pestle of corundum until all particles passed the 40 μ m sieve. About 1 g of the fine powder was applied on a polyethylene sample holder. The X-ray diffractometer (Siemens, D-5000) uses Cu-K_α radiation of wavelength λ =0.154 nm. Diffractograms were recorded from 4-60° 20, in 0.05° 20 increments with 2.5 s counting time per increment.

Fourier-transform infrared spectroscopy (FT-IR)

Solid residues were prepared in the glove-box by separation from the solution with a frit (40 μ m) and rinse them with isopropanol and acetone. Then, the residues were dried for 30 minutes at 40 °C and used as obtained for the analysis. Solutions were filtered (40 μ m) and sealed airtight in a plastic vial. The samples were placed into the spectrometer (BioRad, FTS 175L) and the absorbance was determined from 400-4000 cm⁻¹. For solutions, a pure water sample was recorded and subtracted from the sample spectra.

Environmental scanning electron microscope and energy-dispersive x-ray spectroscopy (ESEM/EDS)

The environmental scanning electron microscope (Philips, XL 30 ESEM-FEG) equipped with EDS-detector was used to examine single phases and to determine their composition. Since samples do not need to be coated, they can be studied in situ without artifacts. Operating conditions were set to 25-10 kV and 10-3 Torr.

Speciation calculation with PHREEQC

Speciation calculations were performed using PHREEQC [9] to explain experimental results and to predict possible mechanisms. The WATEQF4-database and the supplementary reactions given in Table 1 were used for the modelling. The ion-association approaches according to Debye-Hückel and Davies were used to calculate activity coefficients in order to account for the non-ideality of aqueous solutions. These approaches will become unreliable with increasing ionic strength, so the ioninteraction approach according to Pitzer should be used preferably. However, the required parameters for the ion-ion interactions of interest were not available in the known literature.

3 RESULTS

The initial pH of the saturated $Ca(OH)_2$ solutions at pH 12.5 immediately increased after addition of CH₃COOK or HCOOK. Analysis of the resulting solutions by ICP-OES furthermore revealed an increased Ca concentration (Figure 2). The results of the XRD, FT-IR and ESEM/EDS analysis showed no new precipitation products in the solid residue. The FT-IR analysis of a saturated $Ca(OH)_2$ solution with addition of CH₃COOK (2.5 mol/l) shows the same spectrum as a solution of CH₃COOK (2.5 mol/l) alone. Compared to a solution of Ca(CH₃COO)₂ (2.5 mol/l), there are only minor differences, mostly in intensity of the typical CH₃COO⁻ bands at 1553 cm⁻¹ and 1416 cm⁻¹ (Figure 3). For comparison, the pH in the saturated solution of CaSO₄·2H₂O increased only slightly after addition of CH₃COOK, but the Ca concentration increased considerably (Figure 4).

For a saturated solution of $Ca(OH)_2$ with solid $Ca(OH)_2$ in excess and addition of CH_3COOK , the speciation calculation shows an increasing pH that is strongly related to the formation of aqueous Ca-acetate complexes (Figure 5). At low acetate concentrations $CaCH_3COO^+$ is the dominant complex, at higher concentrations $Ca(CH_3COO)_2$ is more stable. Although the concentration of Ca^{2+} is decreased, the total Ca concentration in the solution increases. With increasing addition of CH_3COOK , the increase of pH and the total Ca concentration slows down. Compared to the total CH_3COO^- concentration (not shown in Figure 5), there are only a few percent of Ca-acetate complexes in the solution. Comparable results were obtained for HCOOK.

The ICP-OES analysis for the pore solutions of the vacuum-immersed cement paste samples showed an increase of the K concentration for the samples immersed in the deicer solutions (Figure 6). However, the pH was not as much affected than expected and even decreased for the highalkali cement pastes (Figure 7). The Ca concentration is always higher in the cement pastes immersed in the deicer solutions than in water. At later ages, the Ca concentration increased especially in the low-alkali cement paste samples immersed in the CH₃COOK-based deicer solution (Figure 8). The SO₄²⁻ concentration (not shown) increased for both cement pastes treated with deicer solution, especially in case of HCOOK. The XRD-analysis (not shown) has revealed a decrease of the ettringite peak for the cement pastes exposed to the deicer solutions.

4 **DISCUSSION**

CH₃COOK (2530 g/l at 20 °C) and HCOOK (3310 g/l at 20 °C) are extremely water-soluble salts of acetic acid (CH₃COOH) and formic acid (HCOOH) respectively. Dissolved in water, the acetate/formate-ions behave according to the acid/base theory of Brønsted-Lowry as a base and absorb protons from water (Equation 1 and 2). Hence, aqueous solutions of CH₃COOK and HCOOK are always slightly alkaline. The pH in such solutions can be calculated by using Equation 3. The smaller the pK_b value the stronger the base, i.e. the greater its tendency to absorb protons. The equilibrium constants of Equation 1 and 2 indicate that only small amounts of the dissolved acetate and formate ions form acetic and formic acid.

$$CH_3COO^- + K^+ + H_2O \leftrightarrow CH_3COOH(aq) + K^+ + OH^- \log K_b = -9.24$$
 (1)

$$HCOO^{-} + K^{+} + H_2O \leftrightarrow HCOOH (aq) + K^{+} + OH^{-} \qquad \log K_b = -10.24$$
(2)

(3)

 $pH = 14.0 - 0.5 \cdot (pK_b - \log c)$

with: $pK_b = -\log K_b = basicity constant$

c = concentration of acetate/formate ions in an aqueous solution, [mol/l]

The speciation calculation for the saturated solution of $Ca(OH)_2$ with addition of CH_3COOK fits considerably well to the experimental results (Figure 9). For HCOOK, the mechanism is assumed to be the same, but at higher concentrations there are greater differences between speciation calculation and experimental results. This is probably caused by using ion-association approaches for the calculation of the activity coefficients, because with increasing ionic strength, the results will become progressively unreliable due to increasing ion interactions [10]. Furthermore, underlying uncertainties to the thermodynamic data also affect the calculation. Regardless of that, the formation of the Ca-acetate and Ca-formate complexes provides an fundamental explanation for the observed increase of pH and the total Ca concentration as well (Equation 4, 5, 6 and 7). Since Ca^{2+} is consumed for the formation of these complexes, more solid $Ca(OH)_2$ is dissolved to maintain the equilibrium (Equation 8). Consequently, the total Ca concentration and the pH increase in the solution.

$$CH_{3}COO^{-} + K^{+} + Ca^{2+} + 2OH^{-} \leftrightarrow CaCH_{3}COO^{+} (aq) + K^{+} + 2OH^{-}$$

$$\tag{4}$$

$$2 \operatorname{CH}_3 \operatorname{COO}^- + 2 \operatorname{K}^+ + \operatorname{Ca}^{2+} + 2 \operatorname{OH}^- \leftrightarrow \operatorname{Ca}(\operatorname{CH}_3 \operatorname{COO})_2 (aq) + 2 \operatorname{K}^+ + 2 \operatorname{OH}^-$$
(5)

$$HCOO^{-} + K^{+} + Ca^{2+} + 2 OH^{-} \leftrightarrow CaHCOO^{+} (aq) + K^{+} + 2 OH^{-}$$
(6)

$$2 \operatorname{HCOO}^{-} + 2 \operatorname{K}^{+} + \operatorname{Ca}^{2+} + 2 \operatorname{OH}^{-} \leftrightarrow \operatorname{Ca}(\operatorname{HCOO})_{2} (aq) + 2 \operatorname{K}^{+} + 2 \operatorname{OH}^{-}$$

$$(7)$$

$$Ca(OH)_2(s) \leftrightarrow Ca^{2+} + 2 OH^-$$
(8)

Because of the decreasing solubility of $Ca(OH)_2$ with increasing pH, there will be not enough Ca in the solution for a precipitation of solid $Ca(CH_3COO)_2$. That is supported by the findings of the XRD, FT-IR and ESEM/EDS analysis, where no new precipitation products could be found [11]. However, the direct detection of the Ca-complexes turned out to be more difficult. The FT-IR analysis provided no indication for Ca-complexes in the solutions on the basis of peak shifts or appearance of new bands (Figure 3). But the speciation calculation shows that only small amounts of the Ca-complexes are formed, which may aggravate the detection. Furthermore, FT-IR is appropriate to identify organic ions, but not necessarily to which inorganic ions they are related to. An indirect verification provides the increased solubility of $CaSO_4 \cdot 2H_2O$ in pure water after addition of $CaSO_4 \cdot 2H_2O$ in pure water (2.0 g/l at 20 °C) is comparable to $Ca(OH)_2$ (1.7 g/l at 20 °C), but in the CH₃COOK solution (2.5 mol/l), the solubility of $CaSO_4 \cdot 2H_2O$ increased about 9-times (21.9 g/l at 20 °C). The reason for this is again the formation of aqueous Ca-acetate complexes. But contrary to $Ca(OH)_2$, a much higher Ca concentration is reached, because the pH increases only slightly, caused by the deprotonation of water (Equation 1), and consequently no $Ca(OH)_2$ will precipitate.

In a study from Diamond et al., even a pH of 15.1 was measured after addition of $Ca(OH)_2$ in a CH₃COOK solution of 50 wt%, which is a typical concentration (6.2 mol/l) of commercial runway deicer solutions [12]. The existence of Ca-acetate complexes was confirmed in study by Frantz using Raman-spectroscopy [13]. It was found that in Ca-acetate solutions the concentration of the aqueous species $Ca(CH_3COO)_2$ increased with increasing addition of solid Ca-acetate to the solution and with increasing temperature. The solubility of $Ca(OH)_2$ in acetate solutions was investigated by Seewald et al. at high temperatures and pressures. It was found that the solubility of $Ca(OH)_2$ increased in acetate solutions due to the formation of the strong aqueous complex of $CaCH_3COO^+$ which is accompanied by an pH increase. Only if the Ca-complex formation was considered in the speciation calculations, experimental results correlated with the calculations [14, 15, 16]. Already Taylor suggested, that organic compounds could attack concrete through complexing of Ca^{2+} and consequent dissolution of $Ca(OH)_2$, but also of hydrated silicate and aluminate phases [17]. In a study of Dramé et al. a severe leaching-dissolution effects and altering processes on cement pastes exposed to acetate solutions [18]. Degradation effects and altering processes on cement pastes exposed to acetate solutions were also reported in other studies [19, 20].

Concretes exposed to deicers based on HCOOK during the cyclic climate storage showed in the ESEM/EDS-analysis massive formations of tricarboaluminate (3CaO :Al₂O₃ :3CaCO₃ :32H₂O) in the surface near zone (Figure 9) [21]. Tricarboaluminate, the carbonate analogue of ettringite [17], was identified by the missing sulphur peak but increased carbon peak in the EDS spectrum [8]. This also indicates an interaction between HCOOK deicing solutions and ettringite.

It is well established that acetates and formates are degradable in contact with oxygen, but certain micro organisms are required to start the reactions (Equation 9 and 10). A degradation mechanism could not be found in the laboratory, but in fact takes places in the field. Due to the lower oxygen demand, degradation of formates occurs faster than of acetates. The rate of degradation increases with rising temperature, but high rates are already achieved at low temperatures [22]. It

is unknown so far, if degradation might be a significant mechanism that could affect ASR in the field. However, if degradation products would come into contact with a saturated solution of $Ca(OH)_2$, poorly soluble $CaCO_3$ will precipitate (Equation 11). Consequently, solid $Ca(OH)_2$ is dissolved to maintain equilibrium and the pH would also rise (Equation 8).

$$2 \operatorname{CH}_3 \operatorname{COOK} + 4 \operatorname{O}_2 \to \operatorname{K}_2 \operatorname{CO}_3 + 3 \operatorname{H}_2 \operatorname{O} + 3 \operatorname{CO}_2 \tag{9}$$

$$2 \text{HCOOK} + \text{O}_2 \rightarrow \text{K}_2\text{CO}_3 + \text{H}_2\text{O} + \text{CO}_2$$

$$2 K^{+} + CO_{3}^{2-} + Ca^{2+} + 2 OH^{-} \leftrightarrow 2 K^{+} + 2 OH^{-} + CaCO_{3}(s)$$
(11)

(10)

The cement paste experiments showed some deviating results from the solubility experiments. In fact, the K concentration increases clearly, indicating the ingress of the deicing solution into the cement paste. However, the pH did not increase for the high-alkali cement samples but decreases. At later ages, only a slight increase of the pH was observed for the low-alkali cement pastes. But generally, the pH and the Ca concentration for the samples immersed in the deicer solutions are higher than for the samples immersed in water. A problem with immersion-techniques is leaching, especially when small samples are used, what probably affected the obtained results, especially the decreasing pH. Also the SO_4^{2-} concentration increased for the samples immersed in the deicer solutions, especially in case of HCOOK. That would support the observed decomposition of ettringite by XRD-analysis, which could also be accompanied by the formation of tricarboaluminate, as observed by ESEM/EDS in concretes exposed to HCOOK-deicers during the cyclic climate storage but not found with XRD [21].

The general findings indicate that due to the higher solubility of portlandite in presence of acetate and formate based deicers, more OH⁻ ions are released and consequently the pH increases. This mechanism would clearly promote the attack of reactive aggregates (i.e. silica) and leads to an accelerated ASR in concrete. This also explains why low-alkali cements are not capable of permanently preventing an ASR in concrete exposed to acetate and formate based deicers, if reactive aggregates are used. The alkalis from the deicers alone would not necessarily cause an ASR, as long as the pH would stay low enough. But now, since solid portlandite is also present in concrete with low-alkali cement, more and more OH⁻ ions will be released gradually due to the increased solubility of portlandite in contact with the acetate and formate ions. If reactive aggregates were used, ASR will occur sooner or later. This was confirmed by results from the cyclic climate storage (Figure 11), which is used successfully as ASR performance-test [23].

5 CONCLUSIONS

Experiments with model pore solutions and cement pastes, as well as speciation calculations were performed to explain why deicers based on acetates and formates can initiate and highly accelerate ASR in concrete with reactive aggregates. The findings from the solubility experiments showed an immediate increase of the pH and of the Ca concentration in saturated solutions of Ca(OH)₂ after addition of CH₃COOK and HCOOK respectively, but no new reaction products were found. Only the speciation calculations indicate the formation of strong aqueous complexes of CaCH₃COO⁺, Ca(CH₃COO)₂ and CaHCOO⁺, Ca(HCOO)₂ respectively, resulting in an increased solubility of Ca(OH)₂. Although Ca-complexes could not be found directly by any performed analysis, experiments with CaSO₄ · 2H₂O support the theory of Ca-complex formation. The cement paste experiments showed some different results and new questions raised, but the findings were generally supported. Ultimately, two chemical mechanisms seem to initiate and accelerate the ASR in concrete with reactive aggregates exposed to deicers based on acetates and formates: (1) excess supply of alkalis and (2) release of OH- ions due to the increased solubility of Ca(OH)2. These mechanisms are able to explain the general results found by the cyclic climate storage, including the observation that low-alkali cements may delay the ASR but can not prevent it permanently in concretes with reactive aggregates and exposed to deicers based on acetates and formates.

6 **REFERENCES**

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Phases and Species	Formula	Reaction	$\log K^{(1)}$	Reference
Portlandite	Ca(OH) ₂ (s)	Ca ²⁺ + 2 OH-	-5.19	[9]
Acetic acid	CH3COOH (aq)	H+ + CH ₃ COO-	-4.76	[10]
Potassium acetate	CH3COOK (aq)	K+ + CH ₃ COO-	+0.34	[10, 15]
Calcium monoacetate	CaCH ₃ COO+ (aq)	Ca ²⁺ + CH ₃ COO ⁺	-0.85	[10, 15]
Calcium diacetate	Ca(CH3COO)2 (aq)	Ca ²⁺ + 2 CH ₃ COO ⁻	-1.25	[10, 15]
Formic acid	HCOOH (aq)	H+ + HCOO-	-3.75	[10]
Potassium formate	HCOOK (aq)	K+ + HCOO-	-0.03	[10, 16]
Calcium monoformate	CaHCOO+ (aq)	Ca ²⁺ + HCOO-	-1.40	[10, 16]
Calcium diformate	Ca(HCOO) ₂ (aq)	Ca ²⁺ + 2 HCOO-	-2.24	[10, 16]

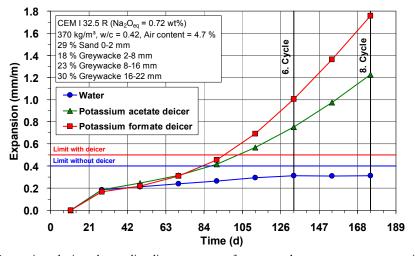


Figure 1: Expansion during the cyclic climate storage for a tested pavement concrete with reactive greywacke, exposed to deicer solutions and water (control)

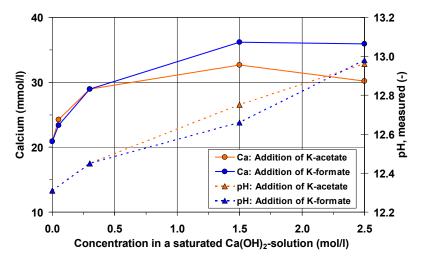


Figure 2: Increase of pH and Ca concentration in a saturated solution of Ca(OH)₂ with addition of CH₃COOK and HCOOK respectively

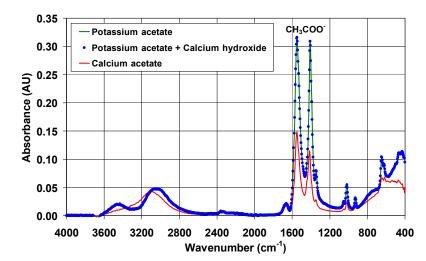


Figure 3: FT-IR spectra (water subtracted) from solutions (2.5 mol/l) of pure $Ca(CH_3COO)_2$, CH₃COOK and a saturated solution of $Ca(OH)_2$ with CH₃COOK

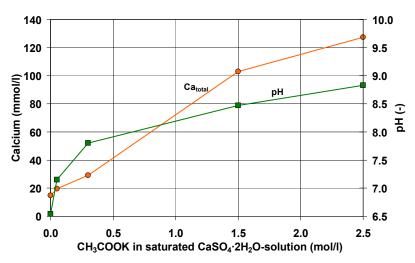


Figure 4: Increase of pH and Ca concentration in a saturated solution of $CaSO_4 \cdot 2H_2O$ with addition of CH_3COOK

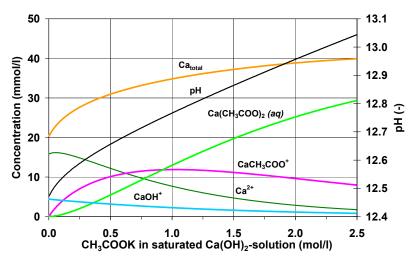


Figure 5: Calculated species distribution in a saturated solution of Ca(OH)_2 with addition of CH_3COOK at 25 $^{\circ}\text{C}$

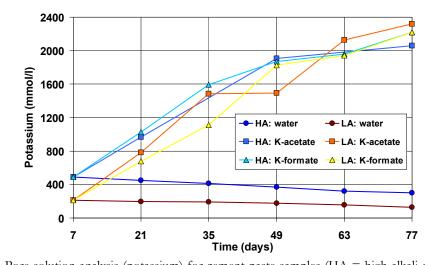


Figure 6: Pore solution analysis (potassium) for cement paste samples (HA = high-alkali cement, LA = low-alkali cement) after vacuum-immersion in water and two commercial deicer solutions

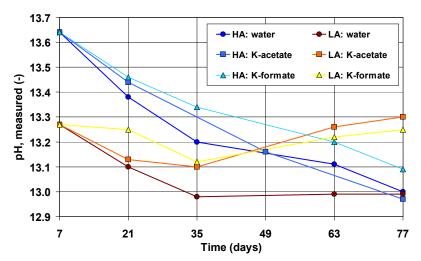


Figure 7: Pore solution analysis (pH) for cement paste samples (HA = high-alkali cement, LA = low-alkali cement) after vacuum-immersion in water and two commercial deicer solutions

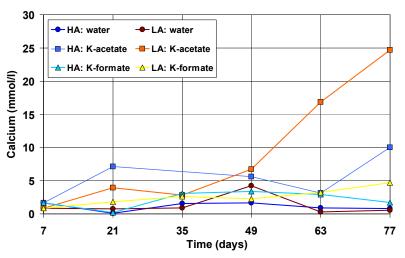


Figure 8: Pore solution analysis (calcium) for cement paste samples (HA = high-alkali cement, LA = low-alkali cement) after vacuum-immersion in water and two commercial deicer solutions

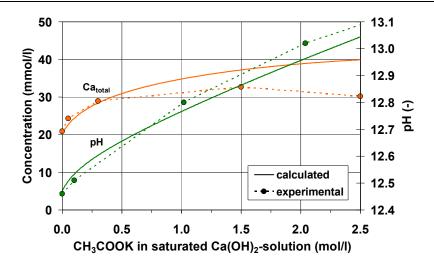


Figure 9: Calculated and measured pH and total Ca concentration in a saturated solution of Ca(OH)_2 with addition of CH_3COOK at 25 $^{\circ}C$

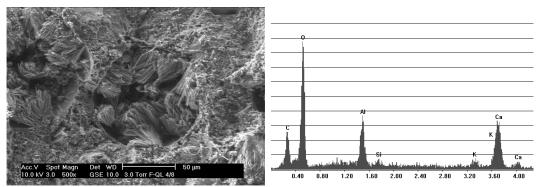


Figure 10: Tricarboaluminate (3CaO·Al₂O₃·3CaCO₃·32H₂O) in concrete air voids near to the concrete surface (3-5 mm) and EDS-spectrum, sample obtained from a concrete exposed to K-formate deicer solution during the cyclic climate storage

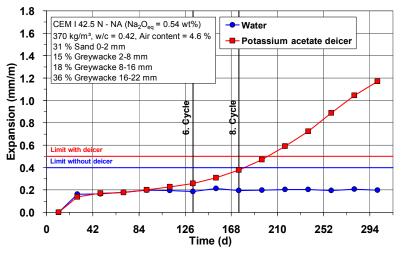


Figure 11: Expansion during the cyclic climate storage for a tested pavement concrete with reactive greywacke and low-alkali cement, exposed to a deicer solution and water (control)